

**Evaluation Manual
for the Authorisation
of plant protection products
according to Regulation (EC) No 1107/2009**

EU part

Plant protection products

**Chapter 7 Ecotoxicology: terrestrial; birds and
mammals**

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ctgb

**Board
for the Authorisation
of plant protection products and biocides**

Chapter 7 Ecotoxicology; terrestrial; birds and mammals

Category: Plant Protection Products

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Changes in the Evaluation Manual

Evaluation manual PPP EU part			
Chapter 7 Ecotoxicology; terrestrial; birds and mammals			
Version	Date	Paragraph	Changes
2.0	January 2014		
2.1	October 2016		Text from data requirements deleted from the Manual, replaced with reference/links to Regulations (EU) No 283/2013 and 284/2013. Short list of data requirements included in the text. General editing and formatting.
		1.3.1	Guidance included on: <ul style="list-style-type: none"> - Uses in playfields and lawns (amateur uses) – presence of voles - PD vole and the use of study by Rinke (1991) - Higher tier refinements agreed upon by the CZSC - Uncertainty analyses - Unclear scenarios - Renewed interception values
2.2	April 2017	1.3.2	- PD vole, updated by the Luthi study - Update higher tier refinements agreed upon by the CZSC (workshop Vienna, 2015)
2.3	May 2018	1.3.2	- Relevant scenarios for herbicidal use in orchards, ornamental/nursery, bush and cane fruit and vineyard - Seedling scenario for treated seeds - PEC sw twa for secondary poisoning (text only)
2.4	January 2020	1.3.1	-Update agreements according to the Pesticide peer review meeting on general recurring issues in ecotoxicology; 185, October 2018
		1	Sentence included on the administrative EFSA guidance
2.5	July 2020		Potato-eating mammals added
2.6	July 2020	1.3.1.2	- Update higher tier refinements agreed upon by the CZSC (workshop Dessau, 2018)

GENERAL INTRODUCTION

This chapter describes the data requirements for estimation of the effects on birds and mammals of a plant protection product and its active substance and how reference values are derived in the EU framework (§1 - §1.5) [Regulation \(EC\) No 1107/2009](#).

1. EU FRAMEWORK

In this document, the procedures for the evaluation and re-evaluation of active substances as laid down in the EU are described; the NL procedure for evaluation of a substance is reverted to when no EU procedure has been laid down. The NL-procedure for the evaluation of a substance is described in §2 - §2.5 of part 2 of the Evaluation Manual (plant protection products). This document aims to give procedures for the approval of active substances and inclusion in [Commission Implementing Regulation \(EU\) No 540/2011](#).

Notifiers preparing an assessment report for active substances need to comply with the relevant guidance, instructions and format laid down in the EFSA [Administrative guidance on submission of dossiers and assessment reports for the peer-review of pesticide active substances](#).

1.1 Introduction

This chapter describes the risk assessment of plant protection products for birds and mammals. The risk of plant protection products to birds and mammals is evaluated to prevent products that present an unacceptable risk to the environment reaching the market.

The [EFSA guidance on risk assessment for birds and mammals \(2009\)](#) is used for the evaluation, at least for all dossiers submitted to the Netherlands since 13/07/2012.

1.2 Data requirements

In order to qualify for inclusion of an active substance in [Commission Implementing Regulation \(EU\) No 540/2011](#) a dossier that meets the provisions laid down in [Commission Regulation \(EU\) No 283/2013](#) and [Commission Regulation \(EU\) No 284/2013](#) of Regulation (EC) No 1107/2009 must be submitted for the active substance as well as for the product.

Generally, EU and OECD guidelines for the protocol of experiments are mentioned in [Commission Communication 2013/C 95/01](#).

When, according to the applicant, a certain study is not necessary, a relevant scientific justification can be provided for the non-submission of the particular study.

1.2.1 Data requirements for the active substance

The data requirements regarding the risk of the active substance to birds and mammals are described in [Commission Regulation \(EU\) No 283/2013](#), point 8.1 (effects on birds and other terrestrial vertebrates).

Point 8.1 consists of the following data requirements:

8.1.1.1: Acute oral toxicity

8.1.1.2: Short-term dietary toxicity to birds

8.1.1.3: Sub-chronic toxicity and reproduction toxicity to birds

It should be noted that if exposure is outside the breeding season this can still lead to effects during breeding seasons. Especially in cases of substances which are expected to have endocrine disruption effects. Thus only the argument that application is outside the breeding season is not considered enough to disregard long-term risk assessment. This was discussed and agreed upon by the experts from the Member states of the Central zone (see 1.3.1.2

below).

- 8.1.2.1: Acute oral toxicity to mammals
- 8.1.2.2: Long-term and reproductive toxicity to mammals
- 8.1.3: Active substance bioconcentration in prey of birds and mammals
- 8.1.4: Effects on terrestrial vertebrate wildlife (birds, mammals, reptiles and amphibians)

Note Ctgb:

There are no guidelines available for investigating toxicity effects of pesticides on amphibians and reptiles. In line with EC 283/2013 this datapoint 8.1.4 is therefore best addressed with available open literature and any other relevant data.

NB: The OECD 231 Amphibian Metamorphosis Assay is a screening assay intended to empirically identify substances which may interfere with the normal function of the HPT axis. It is not intended to quantify or confirm endocrine disruption, or to provide a quantitative risk assessment for amphibians, but only provide evidence that thyroid regulated processes may be sufficiently perturbed to warrant more definitive testing. This assay is therefore not suitable to address datapoint 8.1.4.

- 8.1.5: Endocrine disrupting properties

1.2.2 Data requirements for the product

The data requirements regarding the risk of the plant protection product for birds and mammals are described in [Commission Regulation \(EU\) No 284/2013](#), point 10.1 (Effects on birds and other terrestrial vertebrates).

Point 10.1 consists of the following data requirements:

- 10.1.1.1: Oral toxicity to birds
- 10.1.1.2: Higher tier data on birds
- 10.1.2.1: Acute oral toxicity to mammals
- 10.1.2.2: Higher tier data on mammals

1.2.3 Data requirements for metabolites

Data requirements for metabolites are not clearly reported for the section ecotoxicology. The only reference in [Commission Regulation \(EU\) No 283/2013](#) and [Commission Regulation \(EU\) No 284/2013](#) for ecotoxicology is:

'It may be necessary to conduct separate studies for metabolites, breakdown or reaction products derived from the active substance where non-target organisms may be exposed and where their effects cannot be evaluated by the available results relating to the active substance. Before such studies are performed, the applicant shall take into account the information from Sections 5, 6 and 7.

Studies undertaken shall permit characterisation of metabolites, breakdown or reaction products as being significant or not, and reflect the nature and extent of the effects judged likely to arise.'

Several guidance documents have included a more detailed section on metabolites. More detailed information on data requirements for metabolites is given below, taken from the different guidance documents.

1.2.3.1 *Metabolites* - [Guidance Document on Terrestrial Ecotoxicology \(Sanco/10329/2002 rev 2 final\)](#). As a general principle it should be kept in mind that data requirements for metabolites

mentioned in this section do not always need to be met by means of experimental studies. Applicants may also answer the open questions by means of other available information in support of a scientific and rational risk assessment.

Valuable sources of information are e.g.:

- the molecular structure of the metabolite (is the active part still intact?);
- the presence of metabolites in existing tests with the active substance and/or major metabolites ($\geq 10\%$);
- general knowledge about the relationship between the toxicity of metabolites and the active substances from which these are formed;
- information about the pesticidal activity from biological screening data;
- available knowledge about substances that are related to the metabolites.

No further studies are required where a metabolite is CO₂ or an inorganic substance, not being or containing a heavy metal, or an organic substance with an aliphatic structure, with a chain length of 4 or less, which only consists of C, H, N or O atoms and contains no "structures" or functional groups that are known as ecotoxicologically relevant.

The metabolite is in such cases considered as ecotoxicologically not relevant and has a low risk to the environment.

Generally, a risk assessment is required for all metabolites. Minor metabolites (<10%), however, only need consideration in exceptional cases, e.g. if containing the active moiety of the molecule. By definition the PEC for a minor metabolite is lower than the PEC for the parent compound by more than a factor of 10; accordingly minor metabolites even if up to 10 times as toxic as their parent compound can be considered as safe, provided that the parent compound is safe and also provided that no new concern with regard to persistence is brought in.

Where metabolites are identified in laboratory studies but not in field studies, the field studies must be considered as most relevant unless the difference is caused by the methods used.

Tests with metabolites may not be required in case they are formed relatively rapidly and are present for a short time because they may in such cases have been taken into account in the toxicity tests with the active substance. Such conclusions must, however, be supported by analytical measurements.

Where more than one metabolite is considered as significant, it may be sufficient to conduct tests only with the most important metabolite (highest formation percentage, structure most comparable to the active substance).

Where higher tier studies have been carried out with the active substance, or a relevant formulation, the metabolites may have been taken into account in these studies (depending on the duration of the study and the degradation behaviour of active substance and metabolites).

In principle, the risk assessment for metabolites is the same as for the active substance. In case the metabolite is less toxic than the active substance, it will in most cases entail no greater risk than the active substance, which means that a detailed quantified risk assessment is not required. Exceptions are those metabolites that are more persistent and show more bioaccumulation than the active substance, which may result in differences in long-term exposure.

1.2.3.2 Metabolites - birds and mammals

In the [EFSA guidance on risk assessment for birds and mammals \(2009\)](#) (EFSA GD (2009) a separate section on how the deal with metabolites can be found. In general, metabolites should be taken into account but additional testing on birds or mammals should be prevented a much as possible. See for further information EFSA GD (2009), section 5.4.

1.3 Risk assessment

The risk assessment methodology for birds and mammals has in EU context been elaborated in the [EFSA GD \(2009\)](#).

Each study is summarised and analysed separately. The final conclusion and the endpoint per aspect (such as LD₅₀) are presented in a list of endpoints. Risk assessment is based on comparison with endpoints.

As the EFSA GD (2009) still holds open points, management choices and MS issues, agreements made with different Member States and Dutch views/approaches on certain points are given below.

1.3.1 EU and Zonal agreements

1.3.1.1 EU agreements (*substance*)

Conclusions of the pesticides peer review meeting on general recurring issues in ecotoxicology (Pesticide Peer Review Meeting 133, 23-25 September 2015):

“Mammals” (definition of the relevant reproduction endpoint)

EFSA proposal

- It is proposed to define the ecotoxicologically relevant endpoint for all active substances. This value can then be used in all steps of the risk assessment.
- The definition of an appropriate ecotoxicologically relevant endpoint would limit the need for hazard refinements.
- To define the most appropriate endpoint for the risk assessment it is necessary to derive the lowest endpoint covering all the effects.

Conclusion

Experts expressed concerns that defining an ecotoxicologically relevant endpoint may unnecessarily complicate the assessment i.e. in situations where such endpoint is not needed for the representative use considered at EU level. However, there was a general agreement that the ecotoxicologically relevant endpoint should be identified at EU level in order to ensure consistency in product assessments. The level of conservativeness was also considered and it was concluded that using an ecotoxicologically relevant endpoint in the first tier assessment is unlikely to affect the ability to meet the surrogate protection goal as defined in EFSA (2009).

To facilitate and make the identification of the ecotoxicologically relevant endpoint transparent, it was considered necessary to have advice from the toxicologists over the different parameters investigated in the mammalian toxicology studies. It was agreed to tabulate the necessary information defined in EFSA (2009) (see template in Appendix A of the technical report) and to include this in the ecotoxicology section of the assessment report (DAR/RAR).

Overall, it was concluded to define a single, ecotoxicologically relevant, endpoint and to generate a table (see template in Appendix A of the technical report) summarizing all the relevant information to select such endpoint in collaboration with mammalian toxicologists. The selected endpoint should be used in all the steps of the risk assessment. It was pointed out that there is a need to develop further guidance on the derivation of ecotoxicologically relevant endpoints for mammals.”

This agreement applies to active substance dossiers submitted after September 2015.

[Conclusions of the pesticides peer review meeting on Outcome of the Pesticides Peer Review Meeting on general recurring issues in ecotoxicology - - 2019 - EFSA Supporting Publications - Wiley Online Library:](#)

Extrapolation of studies between different agroclimatic conditions

In relation to the higher tier studies for birds and mammals, the experts considered that the recommendations given by EFSA (2009) are sufficient for spray applications, i.e. any refinements of the risk based on identification of specific focal species and definition of related ecological data should be representative of the area of use of the active substance. This means, for example, to extrapolate a focal species from one zone to another requires consideration of whether the criteria for selecting the focal species are still met. However, the experts noted that higher tier studies for seed treatment uses would need further attention, in order to take into account specific agronomic practices (e.g. sowing rates) and conditions. The experts suggested that any issue related to the agronomic practices maybe addressed in the European Commission's guidance document on seed treatments which is under development and can be considered in the context of the revision of the EFSA Guidance (EFSA, 2009).

Trials for residue decline

Most of the points are explained and discussed in Appendix F –'How to assess and use trials for residue decline in the context of birds and mammals risk assessment' of the conclusions of the pesticides peer review meeting.

Some of the specifics are pointed out below:

-In relation to extrapolation within and between items, the following groups are identified

- 1) dicot plants (green parts and roots)
- 2) monocot plants (green parts and roots)
- 3) fruits
- 4) sprayed seeds (both weed seeds and cereal seeds)
- 5) foliar-dwelling arthropods
- 6) ground dwelling arthropods
- 7) earthworms

Within those groups extrapolation is in general possible. However, it was pointed out that generally common sense and expert judgement should be used, for example, for crop extrapolation within monocot and dicot groups. Extrapolation between the above groups was not considered appropriate. In case of dicot weeds and grass-like weeds, extrapolation is possible for trials performed at a late growth stage. It was also agreed, in general, to avoid extrapolating from maize to grass-like weeds, because maize is a fast-growing crop.

- As regards the dissipation kinetics of a single trial, the recommendations given by FOCUS (2006) were reported. In general, for plants, the single first-order (SFO) kinetic model was recommended. It was agreed that a minimum of five time points should be available for fitting. However, in some exceptional cases four points may be enough (e.g. fast dissipation or metabolites) but there should never be fewer than four time points. Some experts highlighted that the use of pseudoDT50 obtained with the first-order multi-compartment (FOMC) model may be appropriate when SFO cannot be used. For invertebrates, in the case that an attempt to fit a kinetic model provides unreliable DT50 or pseudo DT50, a proposal could be to calculate the time-weighted average factor (fTWA) by integrating the area under the curve (AUC) normalised by the initial value and divided by the averaging period (generally 21 days). The use of the AUC directly in the risk assessment would mean ignoring the residue per unit

dose database given in the EFSA Guidance and this is not recommended. Hence, this kind of approach should only be used to derive an fTWA.

-In relation to the use of degradation kinetics in the risk assessment (e.g. minimum number of trials), the EFSA proposal for plant materials (residue trials performed at least four sites per item and regulatory zone) was agreed although some uncertainties were pointed out. However, it was also agreed that particular climatic conditions of certain areas should be considered to allow extrapolation to some extent (e.g. northern France).

21-day PT

Recently, Ludwigs et al. (2017) published work on how to produce a 21-day time-weighted average value for PT for the wood pigeon. In general, the potential of the approach described in Ludwigs et al. (2017) was recognized, but it was agreed that further consideration on how to use it for risk assessment purposes would be needed. It was further noted that the current approach is based on consumers and the number of consumers is likely to increase with time but the PT for the individual could decrease with time. Therefore, the approach may become less conservative than taking single-day values. For the time being, it was agreed not to use this approach.

These agreements apply to active substance dossiers submitted after June 2019.

1.3.1.2 Zonal agreements

In order to reach agreements on what are the acceptable higher tier refinements for risk assessment on birds and mammals, Ctgb organized the first harmonisation workshop on June 2014 in Wageningen, The Netherlands. The Ctgb notes that higher risk assessment for birds and mammals has become highly complicated and reaching agreement between Member States can be a challenge. 'Ecological refinements' especially often have to be dealt with at Member State level. This is not surprising considering the level of information required to extrapolate 'ecological' studies between countries or zones and was already highlighted in the guidance document (section 6.1.3.2). This means that the EFSA guidance considered ecological refinements in higher tier risk assessment for birds and mammals as Member States specific.

In July and September 2014, the Central Zone Steering Committee (CZSC) discussed the agreements made in this workshop.

In September 2014, all agreements made in the workshop were taken over by the CZSC and were communicated to industry by publishing the agreements on CIRCABC, *except for the agreement on long-term combitox.*

- **Vole – relevance**

The vole needs to be included in the core when relevant in EFSA GD (crop groups in Annex I Table 1.2). Generic refinements should be discussed in the core, MS-specific refinements (related to ecological and agricultural circumstances) in the national addendum.

- **Vole – level of protection**

The risk for voles on population level could be lower than for other mammalian species at the same calculated TER. Population modelling is expected to be a promising way forward to resolve this issue on a scientific basis.

Nevertheless, the majority of MS are not willing to change trigger values for voles.

- **Vole – other issues**

It would also be useful for vole refinement if EFSA would look at RUD for grasses/cereals.

- **PT 1 – %**

Every study should be well described in the core, including presenting both mean and 90th percentile PT values.

- **PT 2 – time period**

PT data should be relevant (or worst-case) for the part of the application period which is associated with the highest risk.

- **PT 3 – Group**

In principle, use consumer only for calculating PT, but additional data can be used in a weight of evidence.

- **PT 4 – LT**

PT (i.e. other numerical parameters) only for long term assessment, unless requirements of EFSA GD 6.1.1 are met.

- **Focal bird species in Central Zone**

It is agreed that a 'living' central zone focal species list is useful for refinement. Species on the list will be automatically accepted as focal species without further data once the list is finalised and agreed. Other species can be used provided that adequate supportive information is submitted by applicants. Endangered species and chick diets are currently not explicitly considered in the risk assessment – this was identified as a research need. [Note Ctgb: to date the central zone focal species list has not been finalised.]

- **MAF**twa***

The majority are in favour of keeping the first tier as it is (however if the applicant already uses the moving time window in the first tier it could also be accepted, because it is worst-case); in the higher tier the moving time window approach should be used. As a default, the 21 day period should be used, unless another period is mentioned in the DAR of the active substance.

Interception is only taken into account at later growth stages with high vegetation coverage (as described in EFSA GD appendix E, table 2). Starting from those stages the FOCUS groundwater interception values can be used for refinement.

- **Pore-water approach**

For now the calculation based on bulk soil concentrations will be used. A calculation based on pore water concentrations would only become meaningful when adequate PEC pore water measurements or calculations are available.

- **Long-term combitox***

There is a need to address combitox effects in long-term risk assessment (Appendix B of [EFSA guidance on risk assessment for birds and mammals \(2009\)](#)).

From July 2014 onwards the notifiers should present both acute and long-term combitox risk assessments in their new core dossiers.

By January 2015 all MS will start evaluating this issue.

The conclusions will be communicated to the Southern and Northern Zone (important for the interzonal dossiers).

* Note that after the meeting additional agreements were reached: A director's consultation group (DCG) has been established within the Central Zone. The DCG encourages harmonisation and cooperation between the Member States. Subjects not agreed upon at the CZSC level are further discussed in the DCG meetings. In the DCG meeting of June 2015 it was decided that the applicants should include the long-term combitox in dossiers submitted from June 1st 2016. The Member States will adopt this approach by July 2016. Please note that the NL has been using this approach since prior to these implementation dates. As the risk assessment addresses the risks associated with exposure of birds and mammals to products, and considering that product data addressing the hazards to birds and mammals are almost never submitted, the combitox risk assessment is a way to address the potential risks from exposure to products.

The long-term combitox should be addressed via the concentration addition (CA) model and should be presented in the dossier for Tier 1. Refinement options and possible consequences are not further outlined. In general, when the CA combitox assessment indicates no acceptable risk, it was agreed that applicants could present a case to demonstrate that adverse effects of the actives are not similar.

- **LD50/10**

All MS consider the LD50/10 in the reproductive risk assessment for birds.

The lowest NOAEL or LD50/10 is used in the risk assessment.

If NOEL from the short-term dietary study is available, this could replace the LD50/10 in TER calculations, when appropriate.

For the higher tier, when a refined NOAEL is available, a comparison with LD50/10 still should be made.

In March 2015, a follow-up workshop was organized by AGES in Vienna, Austria. The topics discussed were as follows: focal bird species, MAF x twa and moving time window, long-term combitox (see above), draft guidance document on measured residues and residue dynamics in plants and arthropods, refinement based on exposure outside the breeding season, avoidance and dehusking factors.

In October 2016 the following decisions were taken by the CZSC:

- **Focal bird species**

In 2016 EFSA launched a project to generate a database including ecological and residue data evaluated in a harmonised way. Though currently it is not clear how this database will and can be used in future, it is not likely to be meaningful to start the work on a comparable database. Before further actions are initiated the final report of the EFSA project will be awaited.

- **Averaging interval for calculation of the twa, MAF x twa or moving time window: discussion of the available tools**

BE and DE have developed tools for calculation of the moving time window. Both tools give similar results and thus any of them can be used in the risk assessment. The application of the time moving window approach was taken note by the CZSC and hence should be applied as explained above (N.B. see the September 2014 agreements) for the second tier risk assessment.

- **Long-term combitox**

The conclusions for the long-term combitox risk assessment for birds and mammals were taken note by the CZSC and hence should be applied as explained above (N.B.

see the September 2014 agreements).

- **Refinement of DT50 and RUD values**
Work on a “guidance document” on how to refine DT₅₀ and RUD values is ongoing.
- **Exposure outside the breeding season: Should exposure outside the breeding season be considered relevant and without the option for waiving the risk, based on delayed effects and/or effects on pair formation and nest building etc.?**
Exposure outside the breeding season can be considered on national level, a respective remark could be included in the core assessment.
- **Refinement parameters for seed treatment: avoidance factor, dehusking factor**
UK provided a proposal how to proceed with studies on de-husking, which was generally agreed by the meeting. Currently, no sufficient information to have a standardised factor for de-husking is available.
Avoidance might be considered by a weight of evidence approach, but not by number in a quantitative way for long-term assessment. It shall also not be used in a quantitative way for acute assessments.
The CZSC did not see the necessity to officially confirm this approach. The incorporation of de-husking and avoidance factors should therefore be followed as discussed above.

Based on ecotoxicology forum discussions in the Central Zone, the following decisions were made by the CZSC in December 2015 regarding secondary poisoning:

Conclusion 1:

PEC to be used in the risk assessment for earthworm-eating birds and mammals:

- Non-persistent substances: take $PEC_{soil, twa, 21 \text{ days}}$ (or $PEC_{soil, max}$ as a worst-case)
- Persistent substances: take $PEC_{soil, twa, 21 \text{ days}} + PEC_{soil, plateau}$ (or $PEC_{soil, max}$ as a worst-case)

Conclusion 2:

PEC to be used in the risk assessment for fish-eating birds and mammals:

- All substances: take RAC-aqua as a screening step, provided that the aquatic risk assessment did not use a twa PEC for one or more groups (and so has a maximum PEC that exceeds the RAC)
RAC-aqua = Regulatory acceptable concentration for aquatic organisms.

Background of these points and their application in the Zonal risk assessment

Conclusion 1:

The [EFSA guidance on risk assessment for birds and mammals \(2009\)](#) (EFSA GD (2009)) presents the possibility of assessing the bioaccumulation potential of lipophilic organic substances (i.e. $\log Pow \geq 3$). In case of the uptake of the substance via the food chain “earthworm to earthworm eating birds and mammals”, two options for assessment are presented, the dry soil and the pore water approach. For the purpose of this document, only the dry soil approach will be discussed.

Regarding the dry soil approach, the EFSA GD (2009), states that the PEC_{soil} with an appropriate TWA according to the reproductive assessment should be used, however the EFSA GD (2009) does not distinguish between persistent and non-persistent substances. Therefore Conclusion 1, provides two possibilities for the assessment. Please note that the environmental fate section should assess if a substance is persistent or not and accordingly calculate the appropriate PEC values. For the ecotoxicological risk assessment, secondary

poisoning of birds and mammals through earthworms, the dry soil approach, Step 3a of the EFSA GD (2009) the residue in earthworms (i.e. PEC earthworm) should be estimated by multiplying the appropriate PEC_{soil} as presented under Conclusion 1 by the BCF_{earthworm}.

Conclusion 2:

For the secondary poisoning of fish eating birds and mammals, the [EFSA guidance on risk assessment for birds and mammals \(2009\)](#) recommends calculating the PEC_{fish} by multiplying the highest PEC water based on the RAC by an appropriate TWA according to the reproductive risk assessment. However it is not clear if by “the relevant PEC_{water}” the EFSA GD (2009) intends that a max PEC or a PEC_{twa 21day} should be used in the assessment.

It was therefore decided that in a screening step (i.e. before first tier) the lowest acceptable surface water concentration for aquatic organisms can be used. This means not using the actual PEC_{sw}, but the critical endpoint with assessment factors (n.b. if the PEC_{sw} is 0.04 mg/L and the lowest endpoint is 50 mg/L with an assessment factor of 100, the 0.5 mg/L can be used for sec poisoning). In this way, the risk is clearly worst case and a TWA does not need to be considered, therefore avoiding discussions on which TWA should be used.

However, this is a screening step, and if the risk is not acceptable then the risk assessment can be conducted with a TWA concentration. Note that it is also possible to skip the screening step, and immediately start with using a twa concentration.

On 20-21 September 2018 the 4th Central Zone Harmonisation Workshop was organised by UBA in Dessau, Germany. The Member States agreed on the following points (see below) and these bullet points were further approved by the CZSC in March 2020.

topic #13: Birds & Mammals - Refinement of RUD and DT50

The MS agreed on the following handling of refinements of RUD and DT50 in the risk assessment for birds and mammals:

- No refinement of RUD and DT50 without detailed justification (by the applicant) would be accepted, and refinements will not be based on one single trial.
- Field trials need replications to be acceptable.
- The trials for refinement need to be close to the GAP, and that deviations need to be justified on a case-by-case basis. If the crop is not eaten, then deviation from the GAP may be accepted.
- All aspects of the study (formulation used, weather conditions, circumstances of application) should be documented in the study report as a comprehensive description of the trials and should also be included in the RR study summary.
- The evaluation of residue dissipation studies for refinement of DT50 should be done in collaboration with fate experts.
- If a refined DT50 value for residues is accepted, the time weighted average concentrations should be determined according to EFSA GD, Appendix H (moving time window approach)
- The zRMS should provide the description and an evaluation on the validity of the submitted studies in the core assessment in order to support a harmonized authorisation on the national level.
- Detailed recommendations are now also available in the EFSA PRAS 185 report and should be followed in addition to the agreements of the 4th CZHW. Generally, the PRAS 185 report overrules the 4th CZHW discussions for contradictory conclusions.

topic #14: Birds & Mammals: Refinement of interception values

The MS agreed that the new interception values according to EFSA guidance on DegT50 values (2014) should be used in the tier 2 assessment, but not yet at tier 1 in zonal assessments.

Please note that the related agreements from September 2014 are also still valid: interception is only taken into account at later growth stages with high vegetation coverage (as described in EFSA GD appendix E, table 2). Starting from those stages the FOCUS groundwater interception values can be used for refinement.

1.3.2 Further elaboration on the risk assessment

Although some agreements have been made in the zonal ecotoxicology workshops, several approaches in higher tier risk assessment have either not been agreed upon, or not discussed at all yet. The Ctgb approach on certain aspects of the higher tier risk assessment is given below.

Adapted interception values

In the EFSA GD (2009), appendix E, interception values relevant for birds and mammals risk assessment are reported. These are based on the FOCUS surface water (sw) and FOCUS groundwater (gw) reports published in 2001 and 2002, respectively and are incorporated in the existing short-cut values. Since the EFSA GD (2009) was published, new surface water and groundwater interception values have become available.

These are [FOCUS groundwater \(May 2014\)](#) and [FOCUS surface water \(May 2015\)](#). Both guidances, which came into force in May 2015 and December 2015 respectively, refer to new interception values published in [EFSA guidance on DegT50 values \(2014\)](#) which came into force in May 2015.

As these guidances are in force, the old interception values as presented in EFSA (2009) should be updated. In the EFSA GD (2009) it is stated that the FOCUS surface water interception values are likely to reflect a worst-case, while the FOCUS gw interception values are more detailed and could be used in higher tier assessment.

As the interception values and short-cut values should be updated, Ctgb proposes using the more detailed FOCUS gw interception values of May 2014 and [EFSA guidance on DegT50 values \(2014\)](#). These documents came into force for both active substance and products in May 2015. Therefore the adapted interception values and short-cut values are valid for dossiers submitted to the Netherlands starting in May 2015. Please note that interception values in early stages are still not accepted, as described in the Appendix E of the EFSA GD 2009 guidance, and agreed upon in the central zone.

The tables below should be used to determine the relevant interception value for a specific growth stage of a crop. The values in the grey cells should not be used, as crop interception is not yet relevant in these stages. Please take care, as in the risk assessment the interception value has to be transformed to a deposition value to determine the residue level on a non-crop food item.

Table 1: Interception (%) by apples, bushberries, citrus and vines dependent on growth stage, as given in Table 1.4 of Appendix C of [EFSA guidance on DegT50 values \(2014\)](#).

Crop	Interception relevant from BBCH	Stage#

	1)					
		<i>BBCH 0-9</i>	<i>BBCH 10-69</i>		<i>BBCH 71-75</i>	<i>BBCH 76-89</i>
Apples ²⁾	≥10	without leaves 50	flowering 60		early fruit development 65	full canopy 65
		<i>BBCH 0-9</i>	<i>BBCH 10-69</i>			<i>BBCH 71-89</i>
Bushberries	≥10	without leaves 40	flowering 60			full foliage 75
		<i>BBCH 0-9</i>	<i>BBCH ≥10</i>			
Citrus	≥10	all stages 80	all stages 80			
		<i>BBCH 0-9</i>	<i>BBCH 11-13</i>	<i>BBCH 14-19</i>	<i>BBCH 53-69</i>	<i>BBCH 71-89</i>
Vines	≥10	without leaves 40	first leaves 50	leaf development 60	flowering 60	ripening 75

Grey cells reflect BBCH stages for which interception cannot be used in the risk assessment.

According to EFSA (2014), Table 1.5, "The BBCH code is indicative (Meier, 2001)". However, Ctgb will take the values in the table unless it is proven by the applicant that they should be different.

¹⁾ This column is added by Ctgb based on Appendix E of EFSA guidance document birds & mammals 2009.

²⁾ 'Apples' is assumed to cover the whole 'orchard' group of Appendix E, with the exception of citrus

Table 2: Interception (%) by other crops dependent on growth stage, as given in Table 1.5 of Appendix C of EFSA (2014).

Crop	Interception relevant from BBCH ¹⁾	Bare-emergence	Leaf development	Stem elongation	Flowering		Senescence Ripening
		BBCH#	BBCH#	BBCH#	BBCH#	BBCH#	BBCH#
		0-09	10-19	20-39	40-89	90-99	
Beans (field + vegetable)	≥50	0	25	40	BBCH 40-49 70	BBCH 50-89 70	80
Cabbage	≥50	0	25	40	BBCH 40-49 70	BBCH 50-89 70	90
Carrots ²⁾	≥40	0	25	60	80		80
Cotton	≥50	0	30	60	BBCH 40-49 75	BBCH 50-89 75	90
Grass ³⁾	Not applicable	0	40	60	90		90
Linseed	≥30	0	30	BBCH 20-29 60	BBCH 30-39 60	70	90
Maize	≥30	0	25	BBCH 20-29 50	BBCH 30-39 50	75 75	90
Oil seed rape (summer and winter)	≥30	0	40	BBCH 20-29 80	BBCH 30-39 80	75 80	90
Onions ⁴⁾	≥40	0	10	25		40	60
Peas	≥50	0	35	55	BBCH 40-49	BBCH 50-89	85

					85	85		
Potatoes	≥40	0	15	60	85			50
Soybean	≥50	0	35	55	BBCH 40-49	BBCH 50-89		65
					85	85		
Spring cereals	≥30	0	0	BBCH 20-29*	BBCH 30-39*	BBCH 40-69*	BBCH 70-89*	80
				20	80	90	80	
Strawberries	≥40	0	30	50		60		60
Sugar beets	≥40	0	20	70 (rosette)		90		90
Sunflower	≥30	0	20	BBCH 20-29	BBCH 30-39	75		90
				50	50			
Tobacco ⁵⁾	Not applicable?	0	50	70		90		90
Tomatoes ⁶⁾	≥50	0	50	70		BBCH 40-49	BBCH 50-89	50
						80	80	
Winter cereals	≥30	0	0	BBCH 20-29*	BBCH 30-39*	BBCH 40-69*	BBCH 70-89*	80
				20	80	90	80	

Grey cells reflect BBCH stages for which interception cannot be used in the risk assessment.

According to EFSA (2014), Table 1.5, "The BBCH code is indicative (Meier, 2001)". However, Ctgb will take the values in the table unless it is proven by the applicant that they should be different.

1) This column is added by Ctgb based on Appendix E of EFSA guidance document birds & mammals 2009.

2) Carrots falls in the crop group 'root and stem vegetables' in the EFSA GD b&m (2009), Table 5. The interception value for carrots is considered to be applicable to this whole crop group, i.e. beetroot, carrot, celeriac, horseradish, Jerusalem artichoke, parsnips, parley root, radishes, salsify, Swedes, turnips, celery, kohlrabi, fennel, etc.

3) A value of 90 is used for applications to established turf. However, for bird and mammal risk assessment, interception refinement is not applicable (see Appendix E of EFSA GD)

4) Bulb vegetables in Table 2 of Appendix E. Onions falls in the crop group 'bulbs and onion like crops vegetables' in the EFSA GD b&m (2009), Table 5. The interception value for onions is considered to be applicable to this whole crop group, i.e. bulbs (like tulips etc.), onions, garlic, shallots, etc. Ctgb includes also leek in this group.

5) This crop is not mentioned in Appendix E. It is not a relevant crop in the Netherlands. It is however grown in the EU so EFSA should include this crop in the updated guidance document for birds and mammals. (In the EU, tobacco cultivation represents some 100,000 ha and 60,000 specialist producers. Tobacco is grown in 12 EU countries. The main producers are Italy, Bulgaria, Greece, Spain and Poland, which account for around 85% of the EU tobacco growing area. From: http://ec.europa.eu/agriculture/tobacco/index_en.htm)

* Citation from EFSA 2014: "BBCH-code of 20-29 for tillering and 30-39 for elongation".

6) Tomatoes falls in the crop group 'fruiting vegetables' in the EFSA GD b&m (2009), Table 5. The interception value for tomatoes is considered to be applicable to this whole crop group, i.e. Tomatoes, peppers, chilli peppers, aubergines, cucumber, gherkins, courgettes melons, squashes, watermelons, etc.

For crops for which interception values that did not change, the values from the EFSA 2009 guidance are still applicable. For crops for which interception values increased, the values for the EFSA 2009 guidance can be considered worst-case, however lower interception values as proposed above can be used.

For crops for which interception values have decreased, the adapted interception values should be used in risk assessment. In the table below, the adapted interception values for the crops are given. The values in bold are more worst case. This only concerns the use in bush and cane fruit, orchards and vines. For crops in *italic*, there is no difference in interception between the previously described interception stages anymore.

Ctgb will actively apply the correction factors on the short-cut values shown in bold in the first tier risk assessment for generic focal species feeding (partly) on plant food items. As an example, the generic focal species in Bush & cane fruit at BBCH ≥ 40 are the small herbivorous mammal and the small omnivorous mammal. These currently have shortcut values for mean and 90th percentile RUDs of 21.7 / 40.9 and 2.3 / 5.2, respectively. Starting May 2015, Ctgb will use shortcut values of 28.9 / 54.4 and 3.06 / 6.9, respectively.

We note that deposition is not necessarily relevant for all components of a mixed diet, but in the first tier we will follow EFSA who in the first tier tables of Annex I applied deposition to all components for omnivorous species nevertheless. In higher tier, the use of deposition should be justified for each individual component.

Table 3: deposition and correction factors relevant for first tier risk assessment of omnivores and herbivores

Crop	Relevant principal BBCH growth stages	Deposition factor according to EFSA 2014	Correction factor on short-cut value
Bare soil	Not applicable	-	
Bulb vegetables	≥ 40	0.4	1
<i>Bush and cane fruit</i>	≥ 10	0.4	0.67
	≥ 20	0.4	0.80
	≥ 40	0.4	1.33
Cereals	≥ 30	0.2	0.40
	≥ 40	0.1	0.33
Cotton	≥ 50	0.25	1
Fruiting vegetable	≥ 50	0.2	0.67
Grassland	Not applicable	-	
Hop	≥ 10	-	-
	≥ 20	0.5	1
	≥ 40	0.3	1
Leafy vegetable	≥ 50	0.3	1
Legume forage	≥ 50	0.3	1
Maize	≥ 30	0.5	1
	≥ 40	0.25	1
<i>Oilseed rape</i>	≥ 30	0.2	0.67
	≥ 40	0.2	0.8
<i>Orchards</i>	≥ 10	0.4	0.5
	≥ 20	0.4	0.67
	≥ 40	0.4	1.33
Ornamentals/nursery	≥ 50	0.3	1
Potatoes	≥ 40	0.15	0.50
Pulses	≥ 50	0.3	1
Root and stem vegetables	≥ 40	0.2	0.67
Strawberries	≥ 40	0.4	1
Sugar beet	≥ 40	0.1	0.40
Sunflower	≥ 30	0.5	1
	≥ 40	0.25	1
Vineyard	≥ 10	0.5	0.83
	≥ 20	0.4	0.80
	≥ 40	0.4	1.33

Extrapolation of ecological refinement studies between countries/zones

As stated in the [EFSA guidance on risk assessment for birds and mammals \(2009\)](#), Extrapolation of study results from one MS or zone to another (section 6.1.3.2) should be done with care.

When using field studies it should be clear that the circumstances in which the study was performed are comparable to the Dutch situation. Therefore an argumentation should be presented when extrapolating from studies performed in other countries than the following:

- Belgium
- Denmark
- Germany
- Ireland*
- Luxembourg
- Northern France
- The Netherlands
- The United Kingdom*

* Birds only. For the United Kingdom and Ireland, it is expected that the agricultural and climatic circumstances are similar to those in the Netherlands. However as this is an island, mammalian species composition and population densities might differ substantially from the continent. Therefore care should be taken for extrapolation of focal species.

Note that if further refinement of ecological data (PD/PT) is proposed, it should first be shown that the proposed focal species is indeed appropriate for the Dutch situation.

- For the refinement on DT_{50} used in MAF-calculation and F_{twa} calculation, the geometric mean is preferred, which is in line with FOCUS kinetics.
- When using refined mean RUD values, the worst-case of either the geometric or the arithmetic mean should be used.

Choice of PT to use in long-term risk assessment

- The 90%-tile PT should be used (as discussed in PRAPeR 80). This is based on the following considerations;

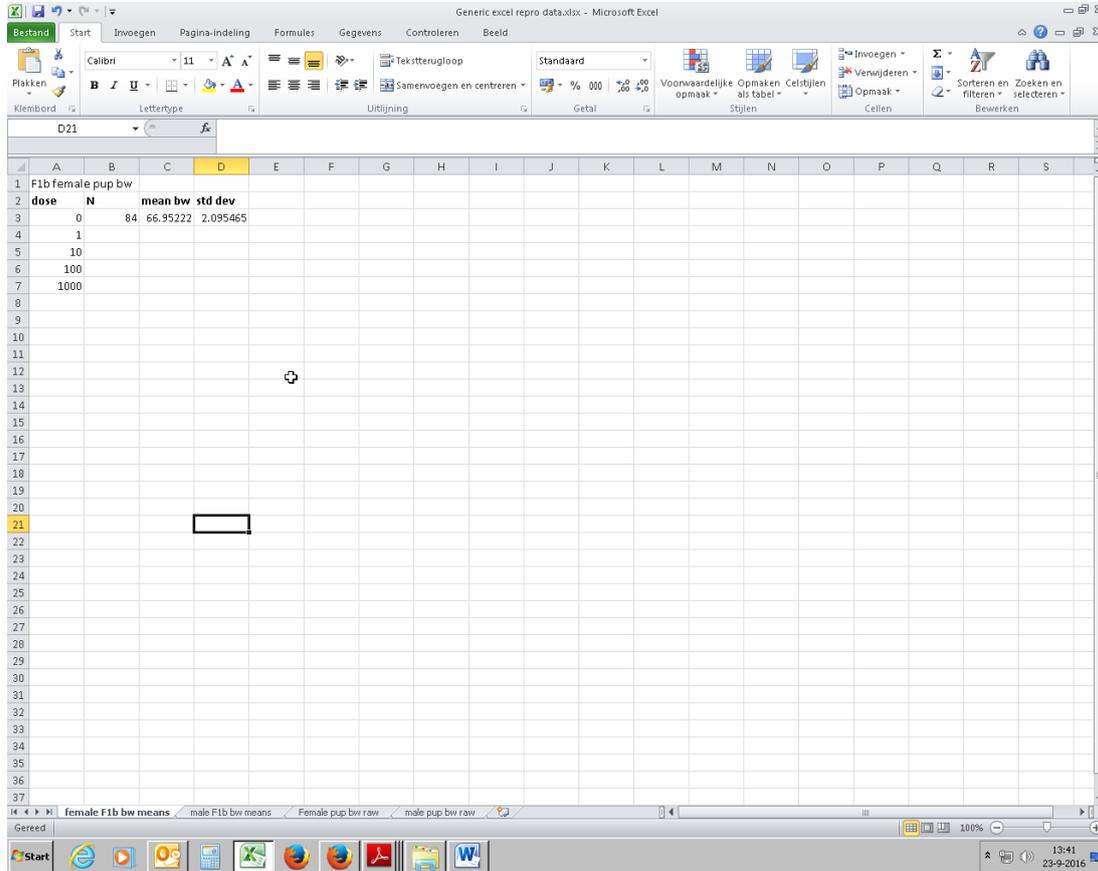
- o Due to uncertainties in deriving PT values (sample size, representativeness of study location, etc) proposed that 90th-tile value should be used. Alternative view that in some cases the mean PT value may be appropriate e.g. depending on sensitivity of focal species.
- o PT is measured over short-term on multiple birds to extrapolate to likely behaviours of individuals over long-term to protect the population using the field.
- o Uncertainty in extrapolating from results from one study in one location to other areas of Europe.
- o Historically, 90th-tile PT values have been used, especially in cases where crops differ to proposed use.

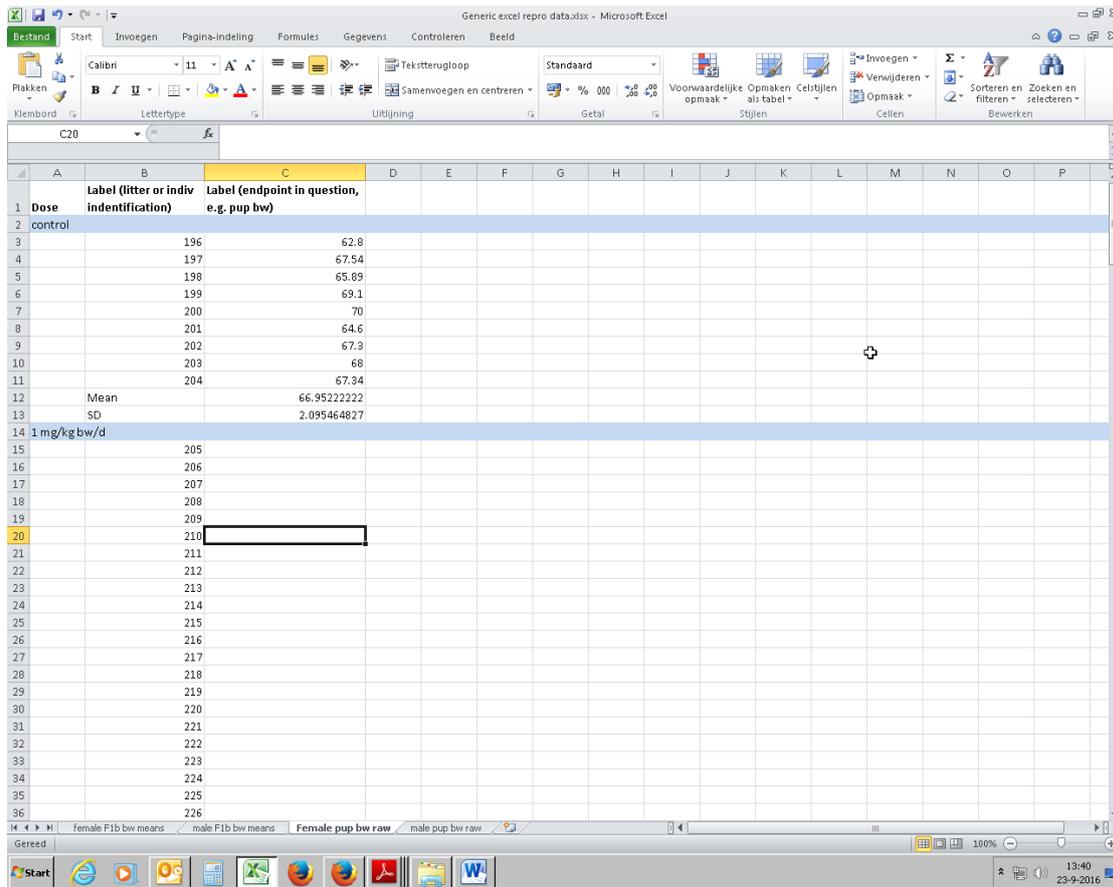
Additionally, the worst-case maximum PT value from field study is used when <10 individuals tracked; when ≥10 individuals tracked use 90th-tile PT value.

Refinement of toxicity endpoint(s)

It is expected that the EU agreed endpoint will lead in most cases, particularly since in many of the active substance renewals greater scrutiny on the endpoints used in risk assessment has become the norm at the EU level. However there may be cases in which an EU endpoint

was not set, or was set provisionally, or where new data has become available. In the event that an endpoint adjustment is requested for either birds or mammals the applicant **must include** the original study report(s) and all relevant raw data in the dossier submission. Additionally, the raw data used for endpoint re-calculation should be presented in an excel file. An example via screen shot(s) is provided, below:





Uncertainty analysis

In the event that higher tier refinements are required in order to achieve an acceptable risk to birds and mammals, the uncertainties involved in consideration of these refinements and the effects of these uncertainties on the conservativeness of the risk assessment should be clearly presented in a concluding uncertainty/weight of evidence table, as shown below (Note: A fictitious uncertainty analysis is included as guidance on the types of things which might be written, but is not exclusive) :

Table B.9.2.3-09 Weight of evidence and uncertainty in the refined risk assessment

Refinement	Source of uncertainty	Discussion and Conclusion regarding uncertainty	Effect on conservativeness
Deposition factor	The use of a DF of 0.3 for the entire period of BBCH ≥ 40 may underestimate the deposition if all applications are made in the later growth stages when the deposition according to FOCUS is 50%.	The tier 1 risk assessment is not overly conservative where deposition is concerned.	-

Refined DT ₅₀	Mainly late BBCH stages were assessed.	Weeds and grass consumed by voles may be at any crop stage at the time of application and at the time of consumption, thus, the trials do not fully represent the breadth of possible growth stages for food items for voles. However, the fact that a majority of the trials were at later growth stages is considered conservative, as growth dilution is less of a factor.	+/-
Default PT	It is assumed that 100% of the diet is consumed in the treated area.	For voles this is a somewhat conservative assumption, however, voles may have relatively small home ranges so it is less conservative than for many other focal species. However, considering the crop structure and the vole's preferred habitat in the grass mat layer, it may be more conservative for certain crops than for others	+
Final Conclusion		The uncertainties involved in the refinements to the risk assessment do not significantly decrease the conservativeness of the assessment.	

Combination toxicity

Combined toxicity should be taken into account. How to deal with potential combinational effects is described in Appendix B of [EFSA guidance on risk assessment for birds and mammals \(2009\)](#). For sub-lethal effects and the effects on reproduction, the EFSA Guidance recommends only performing a risk assessment for combined toxicity on a case-by case basis, because the assessment could be biased by the fact that different effects were the basis of the NOELs (e.g. effects on parent vs effects on offspring), and by the dose spacing in the relevant studies. However, in the central zone it has been agreed that potential combined toxicity should be taken into account for the sub-lethal effects, as well as acute/lethal effects. The issue of combined toxicity is further described in Appendix A of the evaluation manual.

Guidance on certain specific scenarios

Seed treatments/granules/potatoes (tubers), flowerbulbs (dipping)

- in the granule risk assessment for birds (section 5.1.3) a daily grit dose is calculated (DGritD) expressed in dose/ bird/day. As the LD50 or NOEL are expressed in mg/kg bw/d a correction for body weight of the target species should be performed. This is a known flaw in the EFSA guidance and will be corrected in the following update.

- a refinement of the risk evaluation for treated seed is possible by taking the percentage of treated seed that remains at the surface into account. The largest part of the seed is incorporated into the soil and is therefore not accessible for birds and mammals (the digging up of seeds is not taken into consideration because data about this are lacking). As starting point it is assumed that 0.5% of the seeds remains at the surface in case of precision drilling. For standard drilling it is assumed for spring application that 3.3% of the seeds remain at the surface; this percentage is 9.2% in case of autumn application.

The following crops have been studied: onion, sugar beet, maize, alfalfa, flax, pea, spring wheat and winter wheat [Snoo & Luttik 2004](#). It should be determined via expert judgement to which extent other crops are comparable.

-no guidance is given for risk assessment for treated potato tubers. In recent risk assessments, several focal species were proposed:

Birds

The common crane (*Grus grus*) is not a common species in the Netherlands, but it does visit and forage in (parts of) the country when passing it on their migratory route in spring (Mar-Apr) and wintertime (Oct-Dec). For the common crane, a body weight of 5371 g and a daily food intake of 380 g/day are given in addendum 1 to the DAR of flutolanil (October 2006). FIR/bw is 0.071. This value can be used in risk assessment for common crane.

Geese are also known to eat potatoes. However, this normally occurs in wintertime (just before harvest), and not in springtime when treated potatoes will be sown. In conclusion, the common crane is considered to be a reasonable focal species for the Netherlands to indicate possible risks to birds.

Mammals

The risk assessment for effects on mammals is conducted in line with EFSA's Bird and Mammal Guidance Document (EFSA Journal 2009; 7(12):1438). However, the standard scenarios defined by EFSA (2009) do not include treated tubers.

In the EFSA (2008) review report on flutolanil it was pointed out that: "*EFSA recommended that the scenario for granivorous birds and mammals should be revised to take into account 'potato eating' birds and mammals. Badger (Meles meles) was used as relevant focal species*

to represent potatoes-eating mammals.”

The body weight value for the badger would be 10850 g, in line with the value used in the final addendum to the DAR (January, 2008) for flutolanil (from Crocker et al. (2002)¹) and in the re-registration in 2019. In both DARs, an assimilation efficiency of 0.74 is used for vegetable material (from Crocker *et al.* (2002)), which is also the value of fruit for mammals (EFSA, 2009 in Appendix G).

The following values may be used for potatoes: energy content per g of 3.22kJ/g ww and 79.25 % moisture². These are combined with the weight of the badger (10850 g) to calculate a **FIR/bw of 0.132**.

However, in the Netherlands and many other EU countries, the boar is considered the most relevant potato-eating mammal. It is noted that the potato-eating mammal used in the risk assessment for other RARs has been the wild boar (*Sus scrofa*) (used in the EU risk assessment of pencycuron, penflufen and toclofos), based on the adult female wild boar (*Sus scrofa*)³ with a body weight (BW) 60 kg and a daily food intake of 9.98 kg/day (**FIR/bw: 0.17**).

It is noted that due to a lower bodyweight, in the first tier the badger is a more worst-case assumption.

Considering this, in EU and zonal assessments both badger and boar may be used as focal species for potato-eating animals. For applications in the Netherlands only there is no need to show a risk assessment for the badger and the risk assessment can be concluded based only upon the boar.

- degradation/dissipation in seedlings. In the guidance it is stated that the appropriate time window and default degradation and dissipation rates for residues should be considered in the risk assessment via consumption of newly emerged crop shoot. However, the default DT50 of 10 days is only considered valid for surfaces (on leaves and seeds), exposed to weathering. Thus the default DT50 is not automatically considered valid for concentration decline in seedlings bulbs etc.

- Seedling scenario for treated seeds (section 5.2.1 from the EFSA Guidance)_

In determining the DDD for herbivorous birds and mammals following the germination of treated seeds, the approach outlined in section 5.2.1 of the Guidance Document needs to be followed. As stated in the guidance, the seed treatment most closely resembles the “newly sown grassland” or “early post-emergence uses on cereals” scenarios and therefore the relevant focal species are small omnivorous birds and large herbivorous birds and mammals. It is noted however that large herbivorous birds and mammals are mentioned in the text of

¹ Crocker, D., Hart, A., Gurney, J. and McCoy, C. (2002) Project PN0908: methods for estimating daily food intake of wild birds and mammals. Final report. Central Science Laboratory, York.

² USDA National Nutrient Database for standard Reference Release 27: Basic Report 11352, Potatoes, flesh and skin, raw
(<http://ndb.nal.usda.gov/ndb/?format=&count=&max=25&sort=&fg=&man=&facet=&qlookup=potatoes&offset=75>)

³ A bw of 84 kg and a daily food intake of 7 kg/day were used in the EU risk assessments of pencycuron and penflufen (DAR pencycuron, October 2005), however, a body weight of 60 kg and a daily food intake of 9980 g/day were used in the final RAR of toclofos and of flutolanil, which are both more recent.

section 5.2.1, but do not appear in table 19 in the guidance. Therefore when performing a risk assessment, these focal species should be added using for birds a FIR/bw of 0.3 and for mammals a FIR/bw of 0.4 (i.e. goose and rabbit eating cereal shoots resp., as suggested in the guidance).

- quickly dissolving, highly soluble granules. Quickly dissipating, highly soluble granules, for which immediate post-application is prescribed in order to dissolve the granules, a similar situation as for sprayed uses is created. Plant material (i.e. weed seedlings and grass) and arthropods will be contaminated with the product similarly to when a liquid is sprayed. This will result in dietary routes of exposure to birds and mammals and a dietary risk assessment should be performed for these types of formulations using the same scheme as for spray formulations..

Tree nursery crops

This group from the Dutch DTG list is not the same as the ornamental/nursery scenario from the EFSA guidance. Tree nurseries can include more permanent structures, with trees and grass strips in between the rows (i.e. lane trees). In these cases the 'orchard' scenario should be used. This means that when 'tree nursery crops' are not further specified, both ornamental nursery and orchard scenarios are relevant for bird and mammal risk assessment.

Windshields and farmyards

This use is usually for herbicides and is intended for weed removal under trees in a windshield or in a farmyard. For that reason, the scenarios "orchard, not crop directed application all season" can be used for risk assessment purposes for birds and mammals risk assessment.

Herbicide application in orchards, ornamentals/nursery, bush and cane fruit, vineyard

For all herbicide applications (i.e. pre- and post emergence and application on 'zwartstroken') the 'non-crop directed' scenarios apply, in which exposure to food items without interception is relevant. Since this is not consistently addressed in the Guidance, below it is indicated per crop type which scenarios are considered relevant for risk assessment.

Orchards:

For both birds and mammals the risk assessment can be done following Annex I from the Guidance, using the 'Not crop directed application'-scenarios.

Ornamentals/nursery and Bush and cane fruit:

No clear scenarios are specified in the guidance. Use the 'Not crop directed application'-scenarios for orchards for both birds and mammals.

Vineyard:

For birds, no soil directed scenario is specified in the guidance. Use the following generic focal species and short cut values (SV) (SV's without interception):

- Insectivore: Robin (100% soil dwelling invertebrates), SVacute: 7.4, SVchronic: 2.7
- Granivore: Linnet (100% weed seeds), SVacute: 24.7, SVchronic: 11.4
- Omnivore: Lark (25% crop leaves, 25% weed seeds, 50% ground arthropods), SVacute: 24.0, SVchronic 10.9

For mammals, use the 'ground directed application'-scenarios for vineyard from Annex I in the guidance, and in addition include the vineyard scenario for the 'shrew' BBCH 10-19.

'Pre-emergence' uses.

For pre-emergence uses it is often argued that exposure via plants (weeds, grasses) is not

relevant. However, the pre-emergence is usually the stage of the crops of concern, not the stage of the weeds and grasses. The exclusion of the foliar part in the diet can therefore only be used when it is also evident that weeds and grasses are absent at the time of application (for instance, application, directly after tillage). Also for systemic substances, the foliar part of the diet cannot be excluded. Thus for pre-emergence applications, the following additions should be made to the EFSA guidance:

	Presence of weeds and grasses can be excluded (application directly after seeding or planting)	Presence of weeds and grasses cannot be excluded
Systemic	Use the scenarios identified for the next crop stages (BBCH 10-19) to include relevant foliage part of the diet. Use the data for sprayed exposure to foliage, unless measured data of concentrations after systemic exposure in plants is available.	Use the scenarios identified for the next crop stages (BBCH 10-19) to include relevant foliage part of the diet. Use the data for sprayed exposure to foliage, unless measured data of concentrations after systemic exposure in plants is available.
Non-systemic	Foliage is not relevant, bare soil scenario can be used.	Use the scenarios identified for the next crop stages (BBCH 10-19) to include relevant foliage part of the diet.

Risk assessment voles

- Uses in playfields and lawns (amateur uses) – presence of voles

When the application for authorization concerns the use of plant protection products in playing fields and private lawns, it is often believed that the presence of vole will be limited due to the absence of weeds and habitat disturbance due to mowing. It is often suggested that the risk to voles from such applications can be minimized by the application of the restriction sentences on the labels of products.

Another view is that voles will not use the short grasses as habitats or if they use them then it will be for a limited time and therefore a product application will not present a risk. This view is not considered acceptable. In a meeting held in January 2015, the Board agreed that the vole will remain relevant in short grass. The Board considers the vole an too important an indicator species in short grass to stop performing risk assessments for small herbivores in short grass. Additionally, in particular for large lawns, there may be situations where the lawns are no longer mowed or that areas situated in the immediate vicinity of the lawns constitute good habitats for voles. This implies that during daytime/disturbances voles will mainly be present in the undisturbed off-crop area, however, it does not necessarily follow that they will not feed in the treated area once the disturbance is over. It is clear that in case of voles, their presence is influenced mainly by disturbances (i.e. mowing, etc.), rather than the length of grass per se. Voles construct burrows in the thatch/mat layer and feed on the parts of the grass located in that layer (just above the roots and roots). The Ctgb does not accept the presence of dead animals (such as vole, rabbits, and hare) on private lawns and is of opinion that the protection goal as stated in the Regulation 1107/2009 applies to all type of uses. Furthermore, the addition of a warning sentence on the label for non-professional uses is not useful for the following reasons: 1) the NVWA cannot enforce the use of such a sentence for the private uses; 2) the user is unlikely to always read the label; and 3) it is not clear whether this restriction would lower exposure to voles to an acceptable level.

- Uses in grasslands (professional use) - Presence of voles

The group 'grasslands' can include totally different types of grasslands, such as meadows, sod cultivation, golf greens, sport fields etc. Therefore, for the grassland scenario two types of refinements should be conducted: one for applications on grass for seed production and a different one for the application on grass vegetation. For the applications in grass for seed, the Ctgb considers the presence of vole relevant. Voles are known to inhabit the fields used for grass seed production (<http://www.kennisakker.nl/>). Voles are known to graze on the grass seed crops (Hart J.M. et al., EM 9051, September 2012).

Regarding proposed applications in other grass vegetation, the Ctgb agrees that sport fields, golf courses (only fairway, golf greens and golf tees) and the grass sods are intensively managed and the presence of voles can be considered low. The Ctgb has a crop-specific statement regarding the vole: because the grass is kept very short on golf greens and golf tees or grass sods, there is little cover. Therefore, it is not expected that voles occur in these areas in such levels as to constitute a relevant focal species. Instead, the rabbit should be considered as the herbivorous focal species in these areas. In these cases, the scenario of "rabbit" from applications in early cereals is used as a worst-case. For the vole, an off-field risk assessment should be performed, as voles may live in the verges and the less-maintained areas nearby. When performing an off-field risk assessment, a drift level of 10% is assumed, as this off-crop is very close to the in-crop border..

The lawns, playfields, and grass verges will be probably mowed occasionally or even relatively frequently (lawns), however, they are not consistently and rigorously maintained at a low level. Furthermore, if the GAP does not give any information on how frequent these areas should be managed, the Ctgb considers the vole relevant in these cases. This means for professional uses:

Grass for seed – vole is relevant

Grass vegetation (i.e. sport fields, fairway, golf greens and tees) and grass sods – in-field rabbit (take scenario early cereals) and off-field (10% drift) vole

Grass vegetation (i.e. lawns, playfields, and grass verges) – vole is relevant (in-field)

- Vole diet: generic PD refinement

Based on the study by Rinke (1991) "Percentage of volume versus number of species: Availability and intake of grasses and forbs in microtus arvalis. Folia zoologica 40 (2): 143-151" And on Lüthi, M. et al. Nutritional ecology of Microtus arvalis (Pallas, 1779) in sown wild flower fields and quasi-natural habitats. Revue suisse de Zoologie 117 (4): 811-828; december 2010

Studies on the nutrition ecology of *Microtus arvalis* were made by analysing the stomach contents of 363 individuals (186 females and 177 males) caught during 1984-1987 with baited snap traps on five plots of permanent meadow, in central Hessa (Germany). The study investigates vole feeding preferences (mono versus dicot) via stomach content analysis. No exact percentages of each per animal were determined, instead, animals were categorized into 5 potential categories of dicot consumption (20% intervals). Overall, despite the fact that more monocots were available in the surrounding areas (70%), voles showed a preference for dicots, with the majority of voles (all seasons, sexes, ages) showing >80% dicot material in stomach contents.

Table 13 Diet of common voles (% volume) in a meadow in Central Germany (Rinke 1991)

Season	Monocotyledons (% volume)	Dicotyledonos (% volume)	No. of voles
Spring	24	76	23
Summer	25	75	152
Autumn	48	52	188
Total	36	64	363

As this data is based from an extensive study with a high availability of monocots, this can be considered worst-case for situations where the availability of monocots is < 70% of the edible vegetation.

Recently zRMS receive additional data also based on public literature concerning the diet of voles.

In Lüthi et al. 2010 (Nutritional ecology of *Microtus arvalis* (Pallas, 1779) in sown wild flower fields and quasi-natural habitats. *Revue suisse de Zoologie* 117 (4): 811-828; dec. 2010) also an extensive study on the diet of the common vole in monocot and dicot dominated fields was performed.

The study is very detailed (considering that it is public literature) and a large number of samples/voles were considered. Individual measurements or ranges are not reported, therefore differences between individuals cannot be distinguished. Compared with Rinke, the trapping period would seem to favor trapping of juveniles over sub-adults (sub-adults being present in late summer and autumn). This is important because the juveniles were the group in Rinke who actually still ate lots of monocots, particularly in times of year other than summer.

Also, it is not possible to see whether there are differences between the months of the year, as this was not reported. In the sown wild flower areas (SWFF field), vegetation cover was mainly dicot (79%, 81.6% and 79% in the three fields, respectively), while in the quasi natural habitat (QNH) the cover was mainly monocots (82.5, 92.5 and 47.5%). It should be noted that vegetation cover in the third field of the QNH was maximum 64%, while the amount of seeds in the diet is much higher compared to the other two QNH fields with vegetation cover > 100%. However, there is a clear difference in monocot dominated fields (QNH) and dicots dominated fields (SWFF).

Diet results based on quantity analyses, and is given below in the table.

SWFF	field1	field 2	Field 3	average
dicots	16.3	31.8	11.2	19.6
monocots	43.1	36.5	53.3	44.3
seeds	14.8	16.5	27.0	19.4
other (roots)	25.8	15.2	8.5	16.6
QNH				
dicots	17.1	6.2	9.6	11.0
monocots	67.7	81.9	66.0	71.0
seeds	6.6	8.4	17.0	10.7
other (roots)	8.56	3.5	7.4	7.4

Based on the table above, monocot ingestion was on average 53.2% in SWFF, while the rest of the diet was equally divided in dicots and seeds. In QNH, the mean ingestion of monocots was 76.6%, but this could be biased as vegetation cover was low in field 3, and the diet in that field could contain a higher amount of seeds. Again, the rest of the diet is equally divided

between dicots and seeds.

Based on the Rinke study we established an acute PD of 50% monocot and 50% dicot for the acute, and 75% dicot vs 25% monocot for chronic risk assessment, in agricultural fields. The question is if this ratio should change based on this study. Current study does not distinguish between the age of the voles. The way of trapping in this study would suggest a higher amount of juveniles or sub-adults who tend to spend closer to the burrow and take less time searching for preferred food than adults, while the risk assessment for mammals is based on adults. On the other hand, it is not clear how the Rinke study took into account the root and the seed fraction of the diet. All the literature studies on vole diet seem to show that voles do have preferences for certain species, but those preferred species seem to include both monocots and dicots (especially forbs).

For the evaluation manual, as a simplistic refinement, a PD of 50% monocots acute risk assessment is still considered to be appropriate. As the vole is selected as herbivorous species, the other 50% will be selected as dicots, but in the uncertainty analysis, the uncertainty fractions of seeds or roots should be mentioned.

Based on the data of Rinke and Luthi, the following PD refinement can be used in the risk assessment for voles:

		PD	
		RUD unit: non-grass herbs	RUD unit: grass and cereals
Dicot dominated fields (agricultural crops etc.)	Acute	50%	50%
	Chronic	50%	50%
Monocot dominated underground (grasslands, orchards etc.)	Acute	0%	100%
	chronic	25%	75%

1.4 Approval

This section describes the approval criteria for active substances (section 1.4.1) and plant protection products (section 1.4.2 and 1.4.3). For the EU approval procedure of active substances a representative formulation has to be included in the dossier. Therefore section 1.4.1 to 1.4.3 apply. For the zonal applications of plant protection products only section 1.4.2 and 1.4.3 apply.

1.4.1 Approval of the active substance

Annex II of [Regulation \(EC\) No 1107/2009](#) provides the procedure and criteria for the approval of an active substances, safeners and synergists..

Point 3 of Annex II of Regulation (EC) No 1107/2009 gives the criteria for the approval of an active substance.

1.4.2 Evaluation of plant protection products

The principles for the evaluation regarding the effects on the environment are presented in [Commission Regulation \(EU\) No 546/2011](#) (i.e. the Uniform Principles). The specific principles for evaluation for birds and other terrestrial vertebrates are included in Part B Evaluation, point 2.5.2.1.

1.4.3 Decision making for plant protection products

The principles for the decision-making regarding the effects on the environment are presented in [Commission Regulation \(EU\) No 546/2011](#) (i.e. the Uniform Principles).

The specific principles for decision making for birds and other non-target terrestrial vertebrates are included in Part C Decision making, point 2.5.2.1.

Note: the BCF in this case should actually be the BAF (bioaccumulation factor)

1.5 Developments

- The EFSA guidance document should have had an evaluation round in 2012. However this is postponed until 2017.
- Whilst using the EFSA Guidance Document of Risk Assessment for Birds and Mammals (EFSA-Q-2009-00223) several errors in the text have been noticed. These have been collected in a document by FERA and Ctgb refers to this: http://www.hse.gov.uk/pesticides/resources/E/Ecotox_BirdMammal_errors_clarification.pdf
- Agreements on how to deal with DT50 and RUD refinements are discussed on zonal level.
- EFSA is currently working on the 'Data collection for the estimation of ecological data (specific focal species, time spent in the treated areas collecting food, composition of diet) and residues level and residue decline of active substances to be used in risk assessment for birds and mammals' (see EFSA-Q-2015-00211 of Mandate 2015-0078 in the EFSA Register of Questions).